

IMPACTS ON VEGETATION AFTER SELECTIVE CUTTING IN FOREST CONCESSION AREA IN THE SOUTHWESTERN BRAZILIAN AMAZON

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Resumo

Impactos na vegetação após corte seletivo em área de concessão florestal na Amazônia sul-ocidental brasileira. O monitoramento da floresta após as atividades de exploração representa uma ferramenta importante para a gestão florestal, auxiliando no planejamento de uso dos recursos florestais. A presente pesquisa teve por objetivo avaliar os impactos na vegetação arbórea após um ano do término da exploração florestal em área de concessão na Floresta Nacional do Jamari, RO. A vegetação foi avaliada antes da exploração e após a exploração em uma Unidade de Produção Anual (UPA) com taxa de corte de 14,14 m³.ha⁻¹. Delimitaram-se sete parcelas permanentes de 0,5 ha, distribuídas aleatoriamente, onde avaliou-se todos os indivíduos com Diâmetro Altura do Peito (DAP) ≥ 10 cm. A avaliação dos impactos baseou-se em dois critérios, a diversidade e a estrutura da vegetação (fitossociológica, paramétrica, diamétrica e interna). Houve redução de 2% na densidade de plantas e perda de 2 espécies. O estoque volumétrico diminuiu nas classes acima de 80 cm de DAP e a estrutura diamétrica manteve o maior estoque de plantas nas menores classes. As atividades de exploração colaboraram em 3,57 % para aumento da taxa de mortalidade, além de causar danos leves e severos em 15 ind.ha⁻¹. A qualidade do fuste foi, predominantemente, fuste reto (82,36%). Os danos observados estão dentro dos limites esperado para áreas com baixa taxa de corte e submetidas às técnicas de exploração de impacto reduzido. Sugere-se a continuidade do monitoramento a médio e longo prazo afim de descrever os padrões de regeneração e o crescimento das espécies, principalmente as madeireiras, projetando os estoques futuros de corte.

Palavras-chave: Manejo em florestas naturais; Dinâmica florestal; Concessão florestal.

Abstract

The forest monitoring after exploration activities is an important tool for forest management, helping to plan the use of forest resources. This research aimed to evaluate the impacts on tree vegetation after one year of the end of forest harvesting in concession area in the Jamari National Forest, RO. The vegetation was evaluated before and after the exploitation in an Annual Production Unit (APU) with cutting rate of 14.14 m³ ha⁻¹. In total, seven permanent plots of 0.5 ha were divided and randomly distributed, where all individuals with diameter at breast height (DBH) ≥ 10 cm were evaluated. Impact evaluation was based on two criteria, the diversity and structure of vegetation (phytosociological, parametric, diametric, and internal). There was a reduction of 2% in plant density and loss of two species. The volumetric supply decreased in the classes above 80 cm of DBH and the diametric structure maintained the largest supply of plants in the lower classes. The exploration activities collaborated in 3.57% to increase the mortality rate, besides causing slight and severe damages in 15 ind.ha⁻¹. The quality of the stem was predominantly straight (82.36%). The damage observed is within the expected limits for areas with low cut rates and subjected to reduced impact exploration techniques. We suggest the continuous monitoring in the medium- and long- term in order to describe the patterns of regeneration and growth of species, mainly logging, projecting the future supply of cutting.

Keywords: Management in natural forests; Forest dynamics; Forest concession.

INTRODUCTION

The main causes of forest cover loss in the Amazon over the years have been associated, mainly with the opening of federal roads, implementation of settlements, hydroelectric and mining projects, illegal exploitation of high commercial value species, and opening of areas for agricultural activities (VALERIANO *et al.*, 2012).

Thus, the Brazilian government has implemented conservation policies in order to promote the use and protection of natural resources, such as the National System of Nature Conservation Unit (SNUC), which establishes criteria and regulations for creation, implementation, and management of conservation units and the Public Forest Management Regulation, Law No. 11,284/2006, which provides for a management model in which

the government can grant companies and communities the right to manage Conservation Units (CU) for sustainable use, exploiting timber and non-timber resources (BRASIL, 2006).

In this scenario, Rondônia was the first Brazilian state to bid and to approve the concession of forests by the Brazilian Forest Service (SFB) in 2007 in the Jamari National Forest. National Forests are conservation units of sustainable use, with forest cover of predominantly native species and they have as basic objective the multiple use of forest resources and scientific research, with emphasis on methods for sustainable exploitation of native forests (BRASIL, 2000).

In the bidding process, subordinate companies pay for using products and services, and they are subject to meet a series of requirements to the agencies responsible for inspection at the federal and state levels to ensure the sustainable use of natural resources. Thus, the company that won the bid receives the right to practice sustainable forest management for the exploitation of products and services in forest units for periods ranging between 25 and 30 years, applying less harmful practices of exploitation that maintain the forest dynamics after the exploitation of its resources (BRASIL, 2006).

In this sense, because this is a recent process of public forests use, there are few studies assessing the effects of corporate forest management. Some studies addressing conservation units are cited, but focused on an experimental scale (HIGUCHI *et al.*, 2010).

Thus, environmental impacts are understood as any change in physical, chemical, and biological properties of the environment, caused by any human action, directly or indirectly, which may affect the health, safety and well-being of the population, social and economic activities, biome, aesthetic, and sanitary conditions of the environment and the quality of environmental resources (BRASIL, 1986).

According to Souza and Soares (2013) in the evaluation of the post-exploitation impacts, it is necessary to observe the influence of management on richness, composition, structure, and dynamics of the forest and these responses, after exploitation, it can be used as practical information that will assist public and private entities in the management of forests and, consequently, in the elaboration of sustainable forest management plans. Thus, this work aims to evaluate the effect on tree vegetation one year after the end of forest exploitation activities, in a concession area in the Jamari National Forest, Rondônia State (RO).

MATERIAL AND METHODS

LOCATION AND CHARACTERIZATION OF THE STUDY AREA

The study was carried out in an Annual Production Unit (UPA 5) with 1,926.7 ha, located in Forest Management Unit III (UMF III), in Flona Jamari, RO (coordinates 09°00'00" to 09°30'00" S and 62°44'05" to 63°16'54" W). UPA 5 was explored between April and November 2015 at a cutoff rate of 14.14 m³ha⁻¹ (3,042 tree.ha⁻¹).

The region has a climate classified as Tropical Monsoon (Am), hot and humid, with an annual rainfall average of 2300 mm.year⁻¹ and an average annual temperature of 26 °C (ALVARES *et al.*, 2013). The natural vegetation of the region is composed predominantly of Cloud Forest. The relief varies from gently wavy to flat. Latosols and Acrisol are broadly found, with emphasis on occurrence for Red-Yellow Dystrophic Acrisol (MMA, 2005).

Data collection

Data were collected with two forest inventories, which were performed before and after the exploitation activities. In March 2015, seven permanent sampling units were marked and inventoried, with dimensions of 0.5 ha. The sampling units were randomly distributed and subdivided into units of 10 x 10 m, following the methodology described in the guidelines for marking and measuring permanent plots in the Amazon Biome (SILVA *et al.*, 2005). This guideline suggests marking a sampling unit of 0.25 ha for every 250 ha of managed area when the UPA has an area greater than 1000 ha.

Within the sample units, all individuals with diameter at breast height (DBH) ≥ 10 cm were measured and properly numbered with an aluminum platelet. In addition to DBH, a stem identification class (straight, tortuous, inclined, incomplete (≥4.0m), incomplete (<4.0m)) was also collected. Each tree was marked with red ink, at the DBH point, to assist in further re-evaluation.

One year after the end of the exploitation activities, the plots were re-evaluated, assessing all individuals sampled in 2015, plus those who entered in the period.

In this evaluation, in addition to the DBH and the stem quality identification class, the presence of damage (cause, intensity, and position of the damage (SILVA *et al.*, 2005) was also evaluated. The botanical identification was carried out by consultation with the specialist professional granted by the concessionaire company, taxonomists of the Emílio Goeldi Museum, bibliographies (LORENZI, 2002a, 2002b) and specific sites of the area (Flora do Brasil 2020 and Flora brasiliensis). The classification system adopted was APG III.

Data analysis

The maintenance of the integrity of the forest ecosystem, one year after the end of exploitation activities, was evaluated based on two main criteria: effects on the composition of species and effects on tree vegetation structure.

The evaluation of the effects of selective exploitation on the composition of the areas was assessed by the results of floristic richness, Shannon diversity index (H') and Pielou's evenness index (J) (SOUZA and SOARES, 2013). Estimation of the values of the Shannon's and Pielou's indices were generated discarding 255 individuals that were not completely identified.

The effects caused on the structure of the trees were assessed by the results of horizontal or phytosociological structure (density, frequency, dominance, and importance value), parametric structure (supply by DBH class), internal structure (represented by the quality and damage of the stem), mortality and ingress (SOUZA and SOARES, 2013).

The volume of commercial wood of each tree in the sample was estimated to generate the parametric structure and then the volume was grouped by diameter class and transformed per hectare. The individual volume was estimated by the equation of single-entry volume, adjusted for the species exploited in the study area (AMATA, 2014) (Equation 1). Intervals of 10 in 10 cm were considered.

$$V_{c/c} = 0.9494 - 0.0053 * DAP + 0.00086 * DAP^2 \text{ (Equation 1).}$$

Where: $V_{c/c}$ = volume with bark in m^3 ; DBH = diameter at breast height (1.30 m) in cm;

Statistical analysis

The results obtained from the floristic and structural surveys of the forest, before and after exploitation, were submitted to Lilliefors test and Cochran's and Barlett's homoscedasticity. Based on these assumptions, the Student *t*-test was performed, at a 5% probability level, to verify the existence of significant differences between the means of the absolute density (AD), absolute dominance (ADo) and estimated volume (V) relating the year before the exploitation with the year after it.

The analyses and values of the Shannon's and Pielou's indices, as well as the DA, ADo, absolute frequency (AF), and importance value (IV) parameters were obtained with the R CORE TEAM statistical program (2019), using the Agricolae package for statistical tests and the Phyto R for phytosociological structure.

RESULTS

One year after the end of the exploitation activities, a 2% reduction in the total density of individuals with $DBH \geq 10$ cm was observed, from 464.28 ind.ha⁻¹ to 454.86 ind.ha⁻¹ and the disappearance of two genera and two species, however, the number of families was the same (Table 1).

When analyzing the average values of AD, ADo and estimated volume before and after exploration, it was observed that the differences between the means were not significant ($p > 0.05$), thus, note that the total forest supply before and after the exploitation were similar (Table 1).

Table 1. Richness, diversity and middle value of forest of stocks before and after one year the end of logging activities in Annual Production Unit, Flona of Jamari, RO.

Tabela 1. Riqueza, diversidade e valores médios de estoques da floresta antes e após um ano do término das atividade de exploração madeireira em Unidade de Produção Anual, na Flona do Jamari, RO.

Evaluated characteristics	Forest before exploitation	Forest after exploitation	T-test.
Number of families	33	33	-
Number of genera	102	100	-
Number of species	142	140	-
H'*	3.800	3.798	-
J*	0.774	0.775	-
AD (ind.ha ⁻¹)	464.28	454.86	($p > 0.05$)
ADo (m ² .ha ⁻¹)	23.5040	21.8759	($p > 0.05$)
Vc/c (m ³ .ha ⁻¹)	645.84	620.5917	($p > 0.05$)

*Values obtained disregarding unidentified species; Da: absolute density; DoA: absolute dominance and Vc/c: estimated average volume with bark.

The 10 species with the SV also maintained their position in the forest structure after the exploitation. *Protium robustum* was the species with the highest SI before and after exploitation, followed by *Pseudolmedia* spp., *Sclerolobium* spp., *Eschweilera pseudodecolorans*, and *Pouteria torta*, they are being highlighted especially by their higher values of density and frequency in all plots studied (Table 2).

Table 2. Phytosociological structure of the main ten species before and after one year the end of logging activities in Annual Production Unit, Flona of Jamari, RO.

Tabela 2. Estrutura fitossociológica das dez principais espécies, antes e após um ano do término das atividades de exploração madeireira, em unidade de produção anual, na Flona do Jamari/RO.

Species	Before exploitation				After exploitation			
	AD	AF	ADo	IV	AD	AF	ADo	IV
<i>Protium robustum</i> (Swart) D.M. Porter	49.7	100	1.93	20.66	48.6	100	1.87	20.94
<i>Pseudolmedia</i> spp.	40	100	1.29	15.81	38	100	1.25	15.8
<i>Sclerolobium</i> spp.	23.7	100	1.46	13.07	22.9	100	1.4	13.13
<i>Eschweilera pseudodecolorans</i> S.A. Mori	24.6	100	0.85	10.64	24.9	100	0.88	11.24
<i>Pouteria torta</i> (Mart.) Radlk.	26	100	0.72	10.37	24.6	100	0.66	10.17
<i>Beilschmiedia brasiliensis</i> (Kosterm.) Kosterm.	21.1	100	0.66	9.1	20.6	100	0.67	9.31
<i>Peltogyne paniculata</i> Benth.	11.1	100	1.19	9.17	10.9	100	1.06	8.95
<i>Huberodendron swietenoides</i> (Gleason) Ducke	8.3	71.43	0.79	6.38	7.4	71.43	0.78	6.45
<i>Copaifera multijuga</i> Hayne	6	85.71	0.73	5.9	6.3	85.71	0.75	6.31
<i>Apeiba echinata</i> Gaertn.	5.4	100	0.42	4.69	5.4	100	0.43	4.9

Where: AD: absolute density (ind.ha⁻¹); AF: absolute frequency (%); ADo: absolute dominance m².ha⁻¹; IV: importance value (%).

The families with the greatest richness remained. However, two species were lost after the exploitation activities: *Faramea crassifolia* (Rubiaceae) and *Laetia procera* (Salicaceae).

However, when analysing the structure of the species used in forest exploitation, there is a change in the order of importance of the main species, where before exploration were the five most important species in descending order were *Peltogyne paniculata*, *Dinizia excelsa*, *Astronium lecointei*, *Couratari stellata* and *Hymenolobium heterocarpum* and, after forest harvesting, *Peltogyne paniculata*, *Astronium lecointei* and *Couratari stellata* were maintained. *Erismia* cf. *bicolor* and *Qualea paraensis* began to occupy the fourth and fifth position of higher IV, respectively (Table 3).

Table 3. Phytosociological structure before and after one year of logging activities, based on harvested species, in annual production unit, Flona of Jamari, RO.

Table 3. Phytosociological structure before and after one year the end of species splored logging activities of in annual production unit, Flona of Jamari, RO.

Species	Before exploitation				After exploitation			
	AD	AF	ADo	IV	AD	AF	ADo	IV
<i>Peltogyne paniculata</i> Benth.	11.1	100	1.19	9.17	10.9	100	1.06	8.95
<i>Astronium lecointei</i> Ducke	4.9	100	0.34	4.22	4.3	100	0.32	4.13
<i>Couratari stellata</i> A. C. Sm.	4.6	100	0.3	3.99	3.7	85.71	0.26	3.48
<i>Erismia</i> cf. <i>bicolor</i> Ducke	2.3	85.71	0.3	3.25	2.3	85.71	0.31	3.42
<i>Qualea paraensis</i> Ducke	3.4	71.43	0.29	3.22	3.1	71.43	0.3	3.3
<i>Clarisia racemosa</i> Ruiz & Pav.	4	85.71	0.23	3.31	3.4	85.71	0.22	3.25
<i>Diplotropis rodriguesii</i> H.C. Lima	3.4	85.71	0.24	3.23	3.1	71.43	0.24	3.03
<i>Dinizia excelsa</i> Ducke	3.7	71.43	1.38	7.92	2.9	57.14	0.3	2.99
<i>Brosimum rubescens</i> Taub.	4	71.43	0.17	2.8	4	71.43	0.17	2.9

<i>Erisma fuscum</i> Ducke	2.3	85.71	0.18	2.75	2.3	85.71	0.19	2.89
<i>Goupia glabra</i> Aubl	0.6	28.57	0.37	2.17	0.6	28.57	0.4	2.46
<i>Hymenolobium heterocarpum</i> Ducke	2.6	85.71	0.35	3.54	2.3	71.43	0.14	2.41
<i>Caryocar glabrum</i> Pers.	1.4	71.43	0.23	2.53	1.1	57.14	0.15	1.94
<i>Pouteria guianensis</i> Aubl.	1.4	57.14	0.07	1.59	1.4	57.14	0.07	1.64
<i>Cariniana micrantha</i> Ducke	1.4	57.14	0.06	1.57	1.4	57.14	0.06	1.61
<i>Tabebuia incana</i> A.H. Gentry	0.9	42.86	0.02	1	1.1	57.14	0.02	1.34
<i>Bowdichia nitida</i> Spruce ex Benth.	1.1	42.86	0.07	1.28	1.1	42.86	0.07	1.32
<i>Mezilaurus itauba</i> (Meisn.) Taub. ex Mez.	0.9	42.86	0.05	1.16	0.9	42.86	0.06	1.2
<i>Dipteryx odorata</i> (Aublet) Willd.	0.6	28.57	0.09	1	0.6	28.57	0.09	1.03
<i>Simarouba amara</i> Aubl.	0.6	14.29	0.11	0.82	0.6	14.29	0.11	0.86
<i>Peltogyne venosa</i> subsp. <i>densiflora</i> Spruce ex Benth	0.3	14.29	0.01	0.74	0.3	14.29	0.01	0.74
<i>Cedrela fissilis</i> Vell.	0.3	14.29	0.09	0.71	0.3	14.29	0.09	0.74
<i>Caryocar villosum</i> (Aubl.) Pers.	0.6	28.57	0.03	0.73	0.3	14.29	0.02	0.42

Where: AD: absolute density (ind.ha⁻¹); AF: absolute frequency (%); ADo: absolute dominance m².ha⁻¹; IV: importance value (%).

Regarding parametric structure, we found that the AD (ind.ha⁻¹) by diametric class post-logging showed significant changes, while the variables basal area (m².ha⁻¹) and volume (m³ ha⁻¹), generally presented reduction, especially in classes above 80 cm (Table 4). It was verified that, via density, the inverted “J” distribution was maintained, with a higher concentration of individuals in lower classes.

Table 4. Parametric structure, by diameter group, before and after one year the end of logging activities in Annual Production Unit, Flona of Jamari, RO.

Tabela 4. Estrutura paramétrica por classes de diâmetro observada antes e após um ano do término das atividades de exploração madeireira em Unidade de Produção Anual, na Flona Jamari, RO.

DBH class (cm)	AD (ind.ha ⁻¹)		ADo (m ² .ha ⁻¹)		Vc/c (m ³ .ha ⁻¹)	
	Before	After	Before	After	Before	After
10-20	297.1	294.0	4.746	4.742	312.0	309.1
20-30	88.9	85.1	4.129	3.966	118.2	113.3
30-40	36.9	35.7	3.425	3.363	65.8	64.2
40-50	20.9	20.3	3.294	3.213	50.9	49.6
50-60	10.6	10.9	2.401	2.457	33.3	34.1
60-70	4.9	5.1	1.601	1.695	20.5	21.7
70-80	2.0	2.0	0.868	0.887	10.6	10.8
80-90	1.1	0.6	0.652	0.327	7.7	3.9
90-100	0.6	0.6	0.419	0.394	4.8	4.6
>100	1.4	0.9	1.971	0.926	21.9	10.4
Total	464.3	455.1	23.504	21.970	645.8	621.8

Where: AD: Absolute density in individuals per h.ind.ha⁻¹; ADo: Absolute dominance in m² ha⁻¹ and Vc/c: Volume with bark in m³ ha⁻¹.

Out of the 100 dead individuals sampled in the seven plots (28.57 ind.ha⁻¹), 42 died of natural factors (12 ind.ha⁻¹) and 58 of forest management operations (16.57 ind.ha⁻¹), corresponding to 2.58% and 3.57% of the mortality rate, respectively. The size class with the highest number of dead individuals was Class I (10 cm ≤ DBH < 20 cm). The species with the highest number of dead individuals, who had complete identification, were *Protium robustum* and *Sclerolobium* spp. with nine individuals each and *Pouteria torta* with eight. Note that 33 plants that died in the period did not have the identification completed.

The ingress rate was 4.3 %, in Class I ($10 \text{ cm} \leq \text{DBH} < 20 \text{ cm}$). The species with the highest number of ingressing individuals were *Sclerolobium* spp. (six individuals), *Protium robustum* (five individuals), *Tachigali chrysophylla*, and *Virola* spp. (four individuals).

The damage analysis of the internal structure of the forest indicated that 31.57 ind.ha⁻¹ were affected by the exploitation. Out of these, 15.15 ind.ha⁻¹ presented some type of damage and 16.57 ind.ha⁻¹ were killed. For damage due to natural causes, 22.86 ind.ha⁻¹ were recorded (Table 5).

Table 5. Damage causes (DC) and damage intensity (DI) observed in tree vegetation after a year of the end of the activities of forest exploration in Annual Production Unit, Flona Jamari, RO.

Tabela 2. Causa (CD) e Intensidades de danos (ID) observados na vegetação arbórea após um ano do término das atividades de exploração florestal em Unidade de Produção Anual, na Flona Jamari, RO.

DC/DI	Mild (ind.ha ⁻¹)	Severe (ind.ha ⁻¹)	Dead (ind.ha ⁻¹)	Total
Natural damage	6.00	16.86	-	22.86
Exploitation-related damage	7.43	7.72	-	15.15
Dead after exploitation	-	-	16.57	16.57
Natural Death	-	-	12.00	12.00
Total	13.43	24.58	28.57	66.58
%	20.2	36.9	42.9	100

Regarding the place of damage caused by management operations, a higher number of individuals with damaged stem was observed, followed by the crown and damaged crown and stem. For natural damage, the highest proportion of damage was observed in the crown and trunk (Table 6).

Table 6. Damage cause (DC) and position of damage (PD) observed in tree vegetation after a year of the end of the activities of forest exploration in Annual Production Unit, Flona Jamari, RO.

Tabela 3. Causa (CD) e Posições de danos (PD) observados na vegetação arbórea após um ano do término das atividades de exploração florestal em Unidade de Produção Anual, Flona Jamari, RO.

DC/PD	Crown (ind.ha ⁻¹)	Stem (ind.ha ⁻¹)	Crown and stem (ind.ha ⁻¹)	Dead (ind.ha ⁻¹)	Overall Total (ind.ha ⁻¹)
Natural damage	15.43	7.43	-	-	22.86
Exploitation-related damage	6.29	8.00	0.86	-	15.15
Dead after exploitation	-	-	-	16.57	16.57
Natural Death	-	-	-	12.00	12.00
Total	21.72	15.43	0.86	28.57	66.58

Regarding the quality of the stem of the remaining individuals in the forest, a predominance of straight stem was observed, with 82.36 % of the total density, and individuals with tortuous and, or inclined and, or incomplete, were 17.64%. Considering the commercial value species, explored in the UPA, 10.23% presented straight stem and 1.32% with some bent and, or inclination, and or incomplete stem (Table 7).

Table 7. Quality of the stem of the remaining individuals after one year of the end of the forestry activities in the Production Unit at Flona Jamari, RO.

Tabela 4. Qualidade do fuste dos indivíduos remanescentes após um ano do término das atividades de exploração florestal em Unidade de Produção na Flona Jamari, RO.

Quality of the stem	Unexploited species		Exploited species		Total %
	AD (ind.ha ⁻¹)	%	AD (ind.ha ⁻¹)	%	
Straight	328.29	72.13	46.57	10.23	82.36
Bent	48.86	10.73	3.43	0.75	11.49
Inclined	8.57	1.88	1.14	0.25	2.13
Incomplete $\geq 4 \text{ m}$	13.71	3.01	1.14	0.25	3.26
Incomplete $< 4 \text{ m}$	3.14	0.69	0.29	0.06	0.75
Overall Total	402.57	88.45	52.57	11.55	100.0

DISCUSSION

As observed in this study, there were no significant differences for the density of tree individuals per hectare, after exploitation activities in UPA 5, in Flona do Jamari. Souza *et al.* (2017) verified a similar short-term scenario in a management area in dense Cloud Forest submitted to a exploitation rate of $20.02 \text{ m}^3 \cdot \text{ha}^{-1}$, in the state of Amazonas.

As a consequence of the small reduction in the number of individuals after activities, changes in forest richness were found only at the level of genus and species, that is, two species and two genera were eliminated. However, these changes did not affect the diversity indices, which showed the same trend before and after exploitation. This scenario was also observed by Lima *et al.* (2018) in an exploitation area at Flona do Tapajós, one year after harvesting.

These results are significant indicators for the evaluation of policies for the use of forest resources and quality control of forest management, especially in forest concessions, which deal with the use of resources in public forests, where management cannot compromise the quality and efficiency of environmental services, because higher preserved biodiversity in the country are found mostly in Conservation Units.

Thus, the regulation is clear regarding the preservation of forest resources and when they are not fulfilled can lead to the termination of the concession (Brasil, 2006). In this sense, the techniques adopted by the concessionaire for the removal of wood in the area caused few changes in the composition and structure of the analyzed vegetation. There was only a small reduction in density, dominance, and volume, which is expected since the supply undergoes changes due to harvest and mortality. In the area, the low harvest rate adopted ($14.14 \text{ m}^3 \cdot \text{ha}^{-1}$) and the use of reduced effect exploitation techniques, such as opening of fishbone-shaped extensions, directed cutting, among others, contributed to the lower effect on the remaining vegetation.

The species that were lost after exploitation presented only one sampled individual, indicating that groups of species with low density are the most susceptible to exploitation practices.

In this sense, it is highlighted that the concern regarding the exploitation impacts should not only be applied with species that present commercial value, but also the behavior of rare species and other groups of non-arboreal plants, because the integrity of the managed forest must be maintained, as well as its ecosystem. As an example, we mention the work carried out by Magrach *et al.* (2016) who observed, in a forest concession area in Malaysia, the influence of selective cutting on species interaction networks. In this study, the authors observed a significant increase in the composition and density of liana species in exploited areas, which affected the growth rates of the remaining forest.

The species with the highest importance value (*Protium robustum*, *Pseudolmedia* spp., *Sclerolobium* spp., *Eschweilera pseudodecolorans*, and *Pouteria torta*) presented few changes in structure after exploitation. These species, although not characterized as a commercial interest in the region, play an important ecological role in the ecosystem, such as *Protium robustum* (Breu), which widely occur in the Amazon region and presents intense fruiting, serving as food to the fauna (LORENZI, 2002a).

Vieira *et al.* (2015) analyzing the structure and multiple use of tree species in managed forests, in an area of Dense Cloud Forest in Pará, at a cutting rate of $30 \text{ m}^3 \text{ ha}^{-1}$ concluded that after four years of exploitation, significant differences were identified only in total volume and absolute dominance, factors that can affect the order of importance of the species. This was perceived in our study, for the species explored in UPA 5, even with the evaluation in a short period of time it was observed that *D. excelsa* and *H. heterocarpum*, occupied the fourth and fifth place in order of importance, before the exploitation, and they were replaced by *Erisma cf. bicolor* and *Qualea paraenses*, after exploitation. This fact reflects the influence of exploratory activities resulting from the harvest of individuals of these species, which reduced their density and dominance, thus affecting the order of importance in the forest.

Therefore, the monitoring of the remaining individuals of the exploited species requires special attention, in order to evaluate the quality of the forest for a second cutting cycle, which in the studied UPA represented a reduction, especially in the classes above 80 cm of DBH, for density, basal area, and volume, respectively, a class that presented the highest wood removal in the management plan.

Reis *et al.* (2010) observed, after 28 years of exploitation, that forest biomass after exploration in a management area in Flona do Tapajós was recovered. However, the exploited species still presented a negative balance in their supply suggesting a higher cutting cycle, since the species respond differently to forest exploitation, presenting behavior and differentiated regeneration time. However, it is emphasized that in this study the cutting rate was $75.5 \text{ m}^3 \cdot \text{ha}^{-1}$, a value above that one value defined in the legal bases of the govern management plans in the Amazon Biome.

In the study area, one year after the end of the exploitation activities, the forest biomass balance presented negative values, reducing the density of plants above 10 cm of DBH, in the order of $9.09 \text{ ind} \cdot \text{ha}^{-1}$, $1,534 \text{ m}^2 \text{ ha}^{-1}$ of basal area, and $24.04 \text{ m}^3 \cdot \text{ha}^{-1}$ in wood volume, values triggered by the harvest that increases the mortality rate,

exceeding the values of growth and ingress of the forest in the short term. Lima *et al.* (2018) described that the harvesting effect was the main cause of mortality in the exploration area in Flona do Tapajós one year after the end of the activities, resulting in a negative supply balance.

Generally, in areas of short-term forest exploitation, there is a greater supply outflow, because in addition to harvested wood, the mortality rate intensifies due to the increase in individuals who die from exploitation activities. This scenario was described in UPA 5, since the mortality rate due to causes of exploitation represented 3.57% of the total number of individuals who died in the period, and the rate was 2.58% due to natural causes.

However, over time, the mortality rate due to exploitation causes tends to be reduced and ingresses increase, as well as the growth of the remaining species, supported by the opening of clearings (PUIG, 2008), supporting the positive balance in forest biomass. As evidenced by Silva *et al.* (2005) and Reis *et al.* (2010) the forest tends to stabilize approximately five years after exploration, with no significant difference between the understory of an intact forest and that of an exploited forest.

Regarding the analysis of the internal structure of the forest after exploitation, it was found that for every 3.04 trees harvested per hectare, 15.15 ind.ha⁻¹ presented direct damage caused by the exploitation activities, and the severe damage was considered that affecting more individuals (7.71 ind.ha⁻¹) and the stem was the most affected by the damage. Veríssimo *et al.* (2002), analyzing management areas in Paragominas, PA, which were submitted to a cutting rate of 23 m³.ha⁻¹, observed that for every 6.4 trees harvested per hectare, 150 trees ha⁻¹, were severely damaged.

It is also emphasized that 22.86 ind.ha⁻¹ had damage recorded by natural causes, out of these 16.86 ind.ha⁻¹, were characterized with severe damage, considerable value, when compared to the damage directly arising from the harvesting. In this sense, we emphasize the significance of understanding whether part of this damage may be due to the indirect action of exploitation activities, because the opening of the canopy caused by selective cutting can imbalance other trees that are on the edges of the clearings, favoring their fall, or damaging them, in case of winds and, or more intense rains.

Thus, the use of techniques with reduced impact on exploratory activities, as a way to avoid damage to the forest during forest exploitation activities becomes indispensable to provide greater sustainability to forest management, since the damage directly affects the growth of the remaining trees. Vidal *et al.* (2002), studying the growth of trees three years after harvesting in Paragominas, PA, observed that trees without damage grew on average 0.72cm.year⁻¹, those with mild damage grew 0.51 cm.year⁻¹; those with moderate damage grew 0.47 cm year⁻¹ and those with severe damage, grew 0.33 cm.year⁻¹. Thus, according to the authors, the higher incidence of damage may imply a longer period until a second harvest in the area.

Gatti *et al.* (2015) comparing the structures and composition of primary forests, selectively exploited forests, and shallow cutting areas in Africa, noted that exploitation can significantly reduce the biomass of a tropical forest for many decades, reducing the capacity of carbon storage above ground, and creating opportunities for aggressive species to establish themselves and to delay ecological succession, which creates uncertainties regarding the sustainability of selective exploitation.

Regarding the quality of the stem, data indicate that the remaining individuals of commercial interesting species showed little contribution to the percentage value of straight stem (10.23% of the total trees with straight stem). Souza e Souza (2004) performing structural analysis in a Dense Cloud Forest of Terra Firme not explored in the Amazon, found that on average 89% of the individuals in the forest had good stem quality (80 to 100 % utilization) and out of these 34% were represented by trees with commercial value. Therefore, it is noticed that despite the considerable percentage of individuals with adequate stem for commercialization, most of them belong to unexplored species, which implies a lower economic potential of the forest.

CONCLUSIONS

- Short-term forest exploitation activities reduced forest supply, but they did not interfere in floristic composition and species diversity.
- The damage caused to the remaining trees due to the exploitation activities, increased the mortality rate of the forest, occurrence of severe damage and ingress of species with preference to environments with higher incidence of light.
- The management in the conservation unit carried out by the concessionary company proved to be adequate to maintain the quality of the composition, structure, and supplies of the species in the monitored area one year after exploitation, but it is suggested the continuity of monitoring in both medium- and long-term to generate information of the time necessary for complete recovery of damage after harvest, changes in floristic composition, and growth rates of the remaining trees, especially those with commercial value.

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